

VEGETATION DYNAMICS OF ABANDONED CALCAREOUS GRASSLANDS ON RIVER TERRACE SLOPES WITH DIFFERENT ASPECT



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INTRODUCTION

The rate of plant species diversity decrease in semi-natural grasslands after abandonment is very variable. For example, studies of calcareous grasslands in western Europe have shown that 40 to 70 % of plant species disappeared during 10 year abandonment (Willems 1990; 2001), but there was no significant decrease in species richness after 5-10 year abandonment of grasslands in Sweden (Öckinger et al. 2006). Such a great difference is due to a huge number of factors influencing the course of succession. Important factors are management history, edaphic and topographic variables, as well as disturbances, fragmentation etc. (e.g. Ellenberg 1996; Löbel et al. 2006; Eynedl et al. 2008).

Calcareous grasslands are diminishing continuously in Latvia mainly because of land abandonment. However, there is no data about the rate of plant species diversity decrease, but central and western Europe experience is not applicable to Latvian situation in most cases because fragmentation and the impact of intensive agriculture is not so pronounced in Latvia (Rusina 2007).

The aim of the present research was to assess how rapid is the vegetation change and decrease of plant species diversity over seven year period in two calcareous grasslands (dry and mesic) abandoned for nearly 20 years, and to evaluate the impact of topographic and edaphic factors on the rate of changes.

METHODS

Investigations were carried out in two calcareous grasslands (referred to in text as Drubazas and Priednieki, respectively) located in the Abava River valley – the most spectacular calcareous grassland territory in Latvia. Both sites differed in position in relief – Priednieki site was located on the north-east facing slope, but Drubazas site – on the south-west facing slope (Fig. 1), but the vegetation history was the same – grasslands were mowed and grazed for decades and abandoned about 20 years ago with occasional management (grazing in Priednieki and cutting of pines in Drubazas) for two years in late 1990ies.

Vascular plant species and their cover in percentage were recorded each year in July and August from 2001 to 2007 in permanent plots (size of a plot was 1m²) established along transects (parallel to the river) located on the terrace slopes (Fig. 1). Both sites included grassland patches with typical structure and species composition and patches overgrown by expansive grasses and tall herbs (Fig. 1,2,3, Table 2). In total, 20 plots were monitored in Priednieki (10 plots in the most diverse part of the grassland in the *Carex flacca*-*Helictotrichon pratense* community (CHp), 6 plots in a part overgrowing by *Aegopodium podagraria* (*Aegpod*), and 4 plots in a transition from CHp to *Aegpod* community (Tc). 15 plots were monitored in Drubazas (10 plots in the most diverse part of the grassland in the *Filipendula vulgaris*-*Helictotrichon pratense* (FvHp) community, and 5 plots in a part overgrown by *Calamagrostis epigios* (*Calcp*) (Fig. 2,3). Soil samples were taken in 2006. One upper soil layer sample was taken in each community.

RESULTS

Unexpectedly, no one community experienced decrease neither in total species number recorded per community nor in average species number per plot. Vascular plant species richness changed significantly over the research period in both sites: in Drubazas site (SW slope) – species number increased in seven years by 4 species per plot in *Calcp* com. and by 3 species in FvHp com. (Table 1, Fig. 4). The decrease in species diversity (Shannon index) was significant in *Aegpod* community in Priednieki site (Table 1).

Only some species experienced significant changes in abundance. The major increase in abundance was observed for *Aegopodium podagraria* and *Cherophyllum aramicum*, but the major decrease in abundance experienced *Briz media*, *Ononis arvensis*, and *Prunella vulgaris*. In 2007, there were more species with increasing than with decreasing abundance (Table 3). DCA ordination of data collected in 2001 and 2007 revealed that vegetation has changed differently in different communities (Ellenberg indicator values were used to explain environmental gradients). In Drubazas, the *Calcp* plots have moved slightly along axis 1 (nutrient, light and species diversity gradient), but FvHp plots have moved along axis 2 (year and species diversity gradient). In Priednieki site *Aegpod* and Tc plots have moved along axis 1 (increase of moisture and nutrients and decrease of species diversity), and CHp plots have moved along axis 2 (in direction of decrease in moisture and increase of soil reaction) and 3 (year, increase of litter and decrease of herb layer) Table 5, Fig. 7).

Soil physical properties were quite similar for all communities except of CHp community where the proportion of sand was higher (Fig. 5.A). Soil chemical properties are listed in Table 4.

DISCUSSION AND CONCLUSIONS

Calcareous grassland vegetation succession after abandonment in the Abava River valley was slower than reported in many cases in central and western Europe. Even after nearly 20 year abandonment large parts of grasslands still possessed typical species composition and diversity, and during the last seven years increase of aggressive grasses and herbs were observed only in mesic but not in dry grassland.

Grassland on the south-west facing slope was more stable (species number even increased) than NE-facing slope grassland, where encroachment of *Aegopodium podagraria* and *Cherophyllum aramicum* in CHp and especially in Tc community was obvious (CHp and Tc vegetation gradually turn into *Aegpod* vegetation). Although significant decrease in species diversity was not observed it will take plots in future on the north-east facing slope grassland. Topographic and edaphic factors are of main importance to explain differences in vegetation dynamics in both sites. Grassland on the north-east facing slope had soils richer in nutrients and moisture promoting the expansion of nitrophilous herbs. Extreme dryness of the south-west facing slope grassland hindered biomass production and nutrient enrichment of soils in the course of overgrowing.

It is reported that species diversity decreases continuously in abandoned grasslands (e.g. Willems 2001; Luoto et al. 2003; Baur et al. 2006). The current research showed that dry calcareous grasslands can be resistant to deterioration for a long period of time. Several explanations for increase in species diversity on SW-slope grassland for seven years could be put forward:

- research started only 13 years after abandonment, so the main decrease in species numbers could happen before monitoring was launched (unfortunately, the average number of plant species per 1 m² is not known for these communities in Latvia, there are in average 32 species per 9m² in managed FvHp grasslands (Rusina 2007));
- occasional disturbances are very important to keep dry grassland plant communities in dynamic equilibrium: drought periods were natural disturbance and tree cutting and occasional mowing in winter were anthropogenic disturbance in Drubazas site;
- differences in soil properties could explain the fact that there was a slight increase in species number in *Calcp* community and the increase of aggressive grass species (*Calamagrostis epigios* and *Brachypodium pinnatum*) was not observed in FvHp community. Both species are reported as very aggressively encroaching in grasslands shortly after abandonment (Bobbink, Willems, 1987; Rebele, Lehmann 2001; Samodi et al. 2008). The invasion process could be hindered because FvHp community had sick layer of partly weathered dolomites (Fig. 6) and was located higher on the slope promoting more rapid drying out than the soil of *Calcp* community.

Literature

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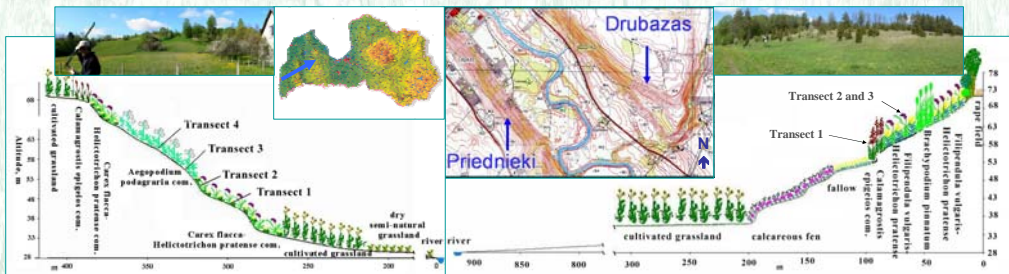


Fig. 1. Location of the transects in a relief profile

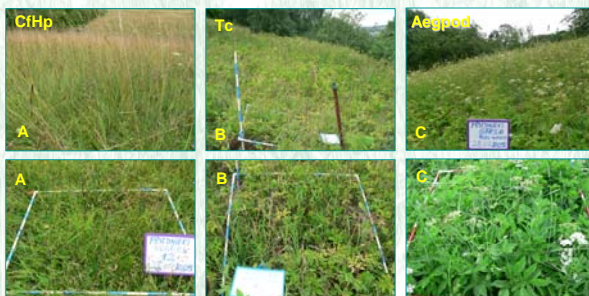


Fig. 2. *Carex flacca*-*Helictotrichon pratense* com.: CHp (A), transitional community: Tc (B), and *Aegopodium podagraria* com.: Aegpod (C).

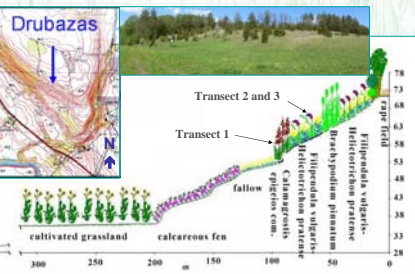


Fig. 3. *Filipendula vulgaris*-*Helictotrichon pratense* com.: FvHp (A) and *Calamagrostis epigios* com.: Calcp (B).

Table 1. Correlation coefficients for seven year observations for species diversity parameters (* p = 0.05; ** p = 0.01).

Parameter	Priednieki		Drubazas	
	CHp (n=10)	Tc (n=4)	Calcp (n=5)	FvHp (n=10)
Year/ total species number per year	0.87**	0.30	0.03	0.84**
Year/ species number per plot	0.64	0.28	-0.12	0.89**
Year/ Evenness	-0.53	-0.79*	-0.54*	0.83*
Year/ Shannon-Wiener index	-0.16	-0.65	-0.90**	0.90**

Table 2. Vegetation table (reduced) with TWINSpan divisions (2001)

Parameter	Priednieki		Drubazas	
	CHp	Aegpod	FvHp	Calcp
pH _{CaCl2} (0-30 cm)	7.1	7.1	6.9	7.1
pH _{CaCl2} (50 cm)	7.6	7.3	7.5	6.9
Bulk density, mg/cm ³	49.6	52.8	43.3	49.7
Organic C %	5.0	5.0	4.7	4.5
N %	2.3	3.2	2.3	2.2
P ₀₋₁₀ mg/100g	2.6	3.8	1.9	2.5
K ₂ O mg/100g	31.1	37.0	5.6	8.4
Ca mg/100g	346.0	215.3	185.5	128.7
Mg ₀₋₁₀ mg/100g	49.0	142.2	107.5	71.9

Table 4. Soil chemical properties

Parameter	Priednieki		Drubazas	
	CHp	Aegpod	FvHp	Calcp
pH	7.1	7.1	6.9	7.1
Organic C %	5.0	5.0	4.7	4.5
N %	2.3	3.2	2.3	2.2
P ₀₋₁₀ mg/100g	2.6	3.8	1.9	2.5
K ₂ O mg/100g	31.1	37.0	5.6	8.4
Ca mg/100g	346.0	215.3	185.5	128.7
Mg ₀₋₁₀ mg/100g	49.0	142.2	107.5	71.9

Fig. 5. Soil fractions



Fig. 6. Soil profile of CHp com. (A), and Aegpod (B) in Priednieki; and FvHp com. (C), and Calcp com. (D) in Drubazas.

Table 5. Correlation coefficients between DCA axes and vegetation and environmental (after Ellenberg indicator values) parameters.

Parameter	Priednieki			Drubazas		
	Axis 1	Axis 2	Axis 3	Axis 1	Axis 2	Axis 3
Year	0.31	-0.21	0.43	0.13	0.44	-0.12
Herb layer	0.31	0.43	-0.43	0.50	-0.48	0.12
Moss layer	-0.05	-0.16	0.28	-0.38	0.27	0.28
Litter layer	0.00	-0.47	0.59	0.67	0.13	-0.13
Species richness	-0.74	-0.33	0.06	-0.59	0.37	0.05
Shannon index	-0.87	-0.23	0.02	-0.30	0.44	-0.09
Evenness	-0.84	-0.19	0.02	-0.09	0.43	-0.14
Light	-0.81	-0.29	0.13	-0.61	0.22	0.08
Temperature	-0.46	-0.21	0.35	0.10	-0.03	0.01
Continentality	0.69	0.17	0.13	0.50	-0.06	-0.08
Moisture	0.71	0.63	-0.11	0.18	0.14	-0.16
Reaction	-0.55	-0.68	0.37	0.19	0.07	-0.08
Nutrients	0.80	0.40	-0.13	0.80	0.04	-0.21

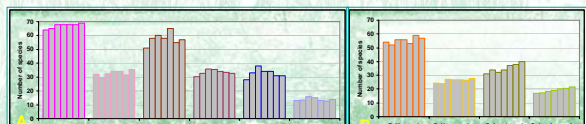


Fig. 4. Changes in species richness from 2001 to 2007 in Priednieki (A) and Drubazas (B).

Table 3. Correlation coefficient between species constancy/cover and observation year (figures in superscript – increase or decrease of cover and constancy from 2001 to 2007 in year). A = cover, C – constancy. * p = 0.05, ** p = 0.01

Species	CHp_C	CHp_A	Tc_C	Tc_A	Aegpod_C	Aegpod_A	Calcp_C	Calcp_A	FvHp_C	FvHp_A
Decreasing species										
<i>Agrostis tenuis</i>	-0.83**	0.89**	-0.11	-0.6	-0.86**	0.86**	a	a	-0.40	-0.40
<i>Briz media</i>	-0.1	-0.82**	-0.11	-0.03	n	n	-0.17	-0.38	0.61	0.81**
<i>Calamagrostis epigios</i>	n	n	n	n	n	n	a	-0.79**	-0.88**	0.05
<i>Cerastium holostoides</i>	-0.79**	-0.79**	-0.31	-0.27	0	-0.41	0	0	0	0
<i>Dactylis glomerata</i>	0.20	-0.76	0.40	-0.46	0	-0.96**	0.6	0.73	0	0.02
<i>Danxia carota</i>	-0.79**	-0.2	-0.82**	-0.79**	n	n	a	a	-0.47	-0.47
<i>Cymodacton conopsea</i>	n	n	n	n	n	n	-0.40	-0.40	-0.75**	-0.58
<i>Linum catharticum</i>	-0.61	-0.61	-0.79**	-0.79**	n	n	-0.40	-0.40	-0.30	-0.30
<i>Potentilla reptans</i>	0	0.75**	-0.61	0.60	-0.90**	-0.72	-0.15	0	0.50	0.50
<i>Prunella vulgaris</i>	-0.85**	-0.94**	-0.89**	-0.87**	n	n	0.62	0.62	-0.17	-0.17
<i>Trifolium pratense</i>	-0.84**	-0.79**	-0.84**	-0.88**	0	-0.43	0.40	-0.40	0	-0.50
<i>Trifolium repens</i>	-0.87**	-0.80**	-0.87**	-0.88**	-0.02	-0.02	a	a	-0.29	-0.29
<i>Vicia cracca</i>	-0.34	-0.42	0.47	0.61	-0.80**	-0.82**	0.61	0.78**	0.58	0.48
Increasing species										
<i>Aegopodium podagraria</i>	0.79**	0.87**	-0.61	-0.86	-0.61	0.82**	0.94**	0.92**	a	a
<i>Altopercus pratensis</i>	n	n	n	n	0.89**	0.89**	n	n	n	n
<i>Carex ornithopoda</i>	0.89**	0.89**	0.14	0.14	n	n	-0.47	-0.47	0.72	0.89**
<i>Campanula rapunculoides</i>	0.61	0.52	n	n	-0.11	-0.92**	0.92**	0.92**	0.58	0.58
<i>Carex coryphylla</i>	0.61	0.61	-0.62	-0.62	n	n	0.85**	0.84**	0.89**	0.84**
<i>Carlina vulgaris</i>	0.61	0.61	n	n	n	n	0.79**	0.78**	-0.21	-0.22
<i>Centaurea jacea</i>	0.82**	0.80**	0.14	0.36	n	n	-0.41	-0.22	0.00	0.68
<i>Cherophyllum aramicum</i>	0.76**	0.83**	0.41	0.75**	0.00	0.95**	n	n	n	n
<i>Deschampsia cespitosa</i>	0.72	0.72	0.00	0.89**	0.79**	0.79**	n	n	n	n
<i>Equisetum arvense</i>	0.55	0.55	0.84**	0.83**	-0.14	0.14	-0.41	-0.41	0.00	0.00
<i>Festuca arundinacea</i>	0.39	0.32	0.89**	0.81**	n	n	-0.61	-0.78**	0.72	0.41
<i>Festuca ovina</i>	-0.41	-0.41	n	n	n	n	n	n	a	0.84**
<i>Fragaria viridis</i>	0.36	0.47	0.69	0.61	n	n	0.78**	0.78**	0.41	0.42
<i>Geranium sylvaticum</i>	0.82**	0.56	0.63	0.55	-0.55	-0.26	n	n	n	n
<i>Glechoma hederacea</i>	n	n	0.43	0.43	0.88**	0.83**	n	n	n	n
<i>Knautia arvensis</i>	0.77**	0.91**	-0.14	-0.14	n	n	0.41	0.42	0.75	0.75
<i>Lysimachia nummularia</i>	0.64	0.66	-0.61	-0.32	0.11	0.87**	n	n	n	n
<i>Pteris hibernica</i>	0.81**	0.66	0.81**	0.86**	n	n	n	n	n	n
<i>Ranunculus acris</i>	-0.56	-0.56	0.78**	0.79**	-0.41	-0.41	0.00	0.00	0.62	0.62
<i>Ranunculus cassubicus</i>	n	n	-0.14	-0.14	0.86**	0.86**	n	n	n	n

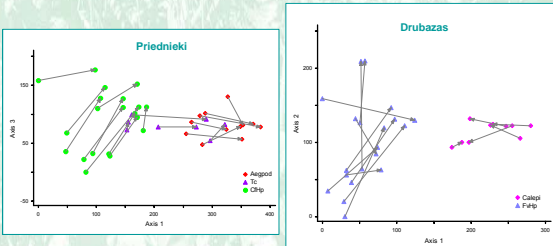


Fig. 7. DCA ordination of plots. Successional vectors join the same plot in 2001 and 2007.

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